

1

COMMENTARY

2

Role of the atmospheric pollution in the Covid-19 outbreak risk in Italy

3

*Daniele Fattorini and Francesco Regoli**

4

Dipartimento di Scienze della Vita e dell'Ambiente (Disva), Università Politecnica delle Marche

5

(Univpm), Via Brecce Bianche, 60100, Ancona, Italy.

6

**Author for correspondence Prof. Francesco Regoli, e-mail f.regoli@staff.univpm.it*

7 **Abstract**

8 After the initial outbreak in China, the diffusion in Italy of SARS-CoV-2 is exhibiting a clear
9 regional trend with Northern areas being the most affected in terms of both frequency and
10 severity of cases. Among multiple factors possibly involved in such geographical differences, a role
11 has been hypothesized for atmospheric pollution. We provide additional evidence on the possible
12 influence of air quality, particularly in terms of chronicity of exposure on the spread viral infection
13 in Italian regions. Actual data on COVID-19 outbreak in Italian provinces and corresponding long-
14 term air quality evaluations, were obtained from Italian and European agencies, elaborated and
15 tested for possible interactions. Our elaborations reveal that, beside concentrations, the chronicity
16 of exposure may influence the anomalous variability of SARS-CoV-2 in Italy. Data on distribution of
17 atmospheric pollutants (NO₂, O₃, PM_{2.5} and PM₁₀) in Italian regions during the last 4 years, days
18 exceeding regulatory limits, and years of the last decade (2010-2019) in which the limits have been
19 exceeded for at least 35 days, confirmed that Northern Italy has been constantly exposed to
20 chronic air pollution. Long-term air-quality data significantly correlated with cases of Covid-19 in
21 up to 71 Italian provinces (updated 27 April 2020) providing further evidence that chronic
22 exposure to atmospheric contamination may represent a favourable context for the spread of the
23 virus. Pro-inflammatory responses and high incidence of respiratory and cardiac affections are well
24 known, while the capability of this coronavirus to bind particulate matters remains to be
25 established. Atmospheric and environmental pollution should be considered as part of an
26 integrated approach for sustainable development, human health protection and prevention of
27 epidemic spreads but in a long-term and chronic perspective, since adoption of mitigation actions
28 during a viral outbreak could be of limited utility.

29

30 **Capsule.** Chronic exposure to air pollutants might have a role in the spread of COVID-19 in Italian
31 regions. Diffusion of Covid-19 in 71 Italian provinces correlated with long-term air-quality data.

32

33 **Keywords:** *COVID-19; Atmospheric Pollution; Chronic exposure; viral diffusion; Italy*

34 **Main text**

35 In December 2019, several pneumonia cases were suddenly observed in the metropolitan
36 city of Wuhan (China), as the result of infection to a novel coronavirus (Li et al., 2020; Wu et al.,
37 2020; Xu et al., 2020). This virus was termed SARS-CoV-2 for its similarity with that responsible of
38 the global epidemic Severe Acute Respiratory Syndrome (SARS) occurred between 2002 and 2003
39 (Xu et al., 2020). Patients affected by SARS-CoV-2 infection often experienced serious
40 complications, including organ failure, septic shock, pulmonary oedema, severe pneumonia and
41 acute respiratory stress syndrome which in several cases were fatal (Chen et al., 2020; Sohrabi et
42 al., 2020). The most severe symptoms, requiring intensive care recovery, were generally observed
43 in older individuals with previous comorbidities, such as cardiovascular, endocrine, digestive and
44 respiratory diseases (Sohrabi et al., 2020; Wang D. et al., 2020). The World Health Organization
45 (WHO) has defined this new syndrome with the acronym COVID-19 for Corona Virus Disease 2019
46 (Sohrabi et al., 2020; WHO, 2020a).

47 The drastic containment measures adopted by Chinese government did not prevent the
48 diffusion of SARS-CoV-2, which in a few weeks has spread globally. Italy was the first country in
49 Europe to be affected by the epidemic COVID-19, with an outbreak even larger than that originally
50 observed in China (Fanelli and Piazza, 2020; Remuzzi and Remuzzi, 2020). Other European
51 countries and United States rapidly registered an exponential growth of clinical cases, leading to
52 restrictions and a global lockdown with evident social and economic repercussions (Cohen and
53 Kupferschmidt, 2020; ECDC, 2020). The WHO has recently declared the pandemic state of COVID-
54 19 with over 2.8 million of cases reported and over 201.000 victims worldwide (Cucinotta and
55 Vanelli, 2020; WHO, 2020b; ECDC, 2020, accessed on 27 April 2020).

56 The ongoing epidemic trend in Italy immediately showed strong regional differences in the
57 spread of infections, with most cases concentrated in the north of the country (Remuzzi and
58 Remuzzi, 2020). The distribution of positive cases reported from February 24th to April 27th is
59 summarized in Figure 1A: some areas of Lombardy and Piedmont clearly exceeded 10.000 cases,
60 e.g. 18.371 at Milan, 12.564 at Brescia, 11.113 at Bergamo, 12199 at Turin (data re-elaborated
61 from the official daily reports of the Department of Civil Protection, ICPD, 2020, accessed on 27
62 April 2020). Also, the relative percentage distribution of the positive test rate (Figure 1B) exhibit
63 higher values in Northern Italy despite a certain uncertainty of data due to the different numbers
64 and frequency of oropharyngeal swabs performed in various regions to test coronavirus positivity;

65 mortality rate ranged from 18% in the most affected, northern regions to less than 5% in the
66 others (Figure 1C). Overall these trends are confirmed from the rates of reported COVID-19 cases
67 and of fatal events, expressed as percentage values normalized to the number of inhabitants for
68 regional populations (Figures 1D and 1E), further corroborating significantly greater effects in
69 Northern Italy, both in terms of number of infections and the severity of cases (mortality).

70 To explain such geographical trend, it was initially assumed that restrictions decided by
71 government authorities after the first outbreak in Lombardy, had contained the effects of the
72 infection, preventing its rapid spread to the rest of the country. Some authors, however, from the
73 clinical course of a large cohort of patients, have concluded that the epidemic coronavirus had
74 been circulating in Italy for several weeks before the first recognized outbreak and the relative
75 adopted containment measures (Cerada et al., 2020). In this respect, the differentiated occurrence
76 of infection cannot be fully explained by the social confinement actions. Our elaboration we did
77 not include the analysis of other possible determinants of incidence and mortality, such as age
78 structure, capacity of the healthcare system, duration of the confinement since some of these
79 data are currently not immediately available. Future studies will be necessary to fill such gap of
80 knowledge addressing at which extent all these factors may have mutually contributed to the
81 diffusion of COVID-19 in Italy. However, the current profile of the viral outbreak in Northern Italy
82 suggests that other factors could be involved in the diffusion of infection and mortality rates.

83 Since the presence of comorbidities appeared determinant for the aetiology and severity of
84 the COVID-19 symptoms (Chen et al., 2020; Wang T. et al., 2020; Wu et al., 2020), the role of
85 atmospheric pollution in contributing to the high levels of SARS-CoV-2 lethality in Northern Italy
86 has been hypothesized (Conticini et al., 2020). Association between short-term exposure to air
87 pollution and COVID-19 infection has been described also for the recent outbreak in China (Zhu et
88 al., 2020). The adverse effects of air pollutants on human health are widely recognized in scientific
89 literature, depending on various susceptibility factors such as age, nutritional status and
90 predisposing conditions (Kampa and Castanas, 2008). Chronic exposure to the atmospheric
91 pollution contributes to increased hospitalizations and mortality, primarily affecting cardiovascular
92 and respiratory systems, causing various diseases and pathologies including cancer (Brunekreef
93 and Holgate, 2002; Kampa and Castanas, 2008). Among air pollutants, the current focus is mainly
94 given on nitrogen dioxide (NO₂), particulate matter (PM_{2.5} and PM₁₀) and ozone (O₃), frequently
95 occurring at elevated concentrations in large areas of the planet.

96 The percentage of European population exposed to levels higher than the regulatory limits
97 is about 7-8% for NO₂, 6-8% for PM_{2.5}, 13-19% for PM₁₀ and 12-29% for O₃ (EEA, 2019). Premature
98 deaths due to acute respiratory diseases from such pollutants are estimated to be over to two
99 million per year worldwide and 45.000 for Italy (Brunekreef and Holgate, 2002; Huang et al., 2016;
100 EAA, 2019; Watts et al., 2019).

101 Here, we are providing additional evidence on the possible influence of air quality on the
102 spread of SARS-CoV-2 in Italian regions. Since the effects of air pollutants on human health not
103 only depend on their concentrations but also, if not especially, on chronicity of exposure, we have
104 elaborated the last four years (from 2016 to 2019, EEA, 2020) of regional distribution of NO₂, PM_{2.5}
105 and PM₁₀ is presented in Figure 2. The highest atmospheric concentrations were clearly
106 distributed in the Northern areas (Piedmont, Lombardy, Veneto and Emilia-Romagna), in addition
107 to urbanized cities, such as Rome and Naples.

108 The long-term condition of population exposure is also revealed by the number of days per
109 year in which the regulatory limits of O₃ and PM₁₀ are exceeded (Figure 3A-B): the critical situation
110 of Northern Italy is reflected by values of up to 80 days of exceedance per year (average of the last
111 three years, EAA, 2019). Worthy to remind, ozone is one of the main precursors for the formation
112 of NO₂, and chronic exposure to this contaminant for almost a quarter of a year is undoubtedly of
113 primary importance. The chronic air pollution in Northern Italy is further represented by the
114 number of years during the last decade (2010-2019) in which the limit value for PM₁₀ (50 µg/m³
115 per day) has been exceeded for at least 35 days (Figure 3C). Once again, these data confirm that
116 the whole Northern area below the Alpine arc has been constantly affected by significantly higher
117 levels of these contaminants.

118 The hypothesis that atmospheric pollution may influence the SARS-CoV-2 outbreak in Italy
119 was also tested from the relationships between the confirmed cases of Covid-19 in up to 71 Italian
120 provinces (updated 27 April 2020) with the corresponding air quality data. The latter were
121 expressed as average concentrations in the last 4 years of NO₂, PM_{2.5} and PM₁₀ (Figure 4A-C) and
122 the number of days exceeding the regulatory limits (averages of the last 3 years) for O₃ and PM₁₀
123 (Figure 4D-E). The always significant correlations provided further evidence on the role that
124 chronic exposure to atmospheric contamination may have as a favourable context for the spread
125 and virulence of the SARS-CoV-2 within a population subjected to a higher incidence of respiratory
126 and cardiac affections.

127 It is well known that exposure to atmospheric contaminants modulate the host's
128 inflammatory response leading to an overexpression of inflammatory cytokines and chemokines
129 (Gouda et al., 2018). Clear effects of Milan winter PM_{2.5} were observed on elevated production of
130 interleukin IL-6 and IL-8 in human bronchial cells (Longhin et al., 2018), and also NO₂ was shown to
131 correlate with IL-6 levels on inflammatory status (Perret et al., 2017). The impairment of
132 respiratory system and chronic disease by air pollution can thus facilitate viral infection in lower
133 tracts (Shinya et al., 2006; van Riel et al., 2006).

134 In addition, various studies have reported a direct relationship between the spread and
135 contagion capacity of some viruses with the atmospheric levels and mobility of air pollutants
136 (Ciencewicki and Jaspers, 2007; Sedlmaier et al., 2009). The avian influenza virus (H5N1) could be
137 transported across long distances by fine dust during Asian storms (Chen et al., 2010), and
138 atmospheric levels of PM_{2.5}, PM₁₀, carbon monoxide, NO₂ and sulphur dioxide were shown to
139 influence the diffusion of the human respiratory syncytial virus in children (Ye et al., 2016), and the
140 daily spread of the measles virus in China (Chen et al., 2017; Peng et al., 2020).

141 These evidences and presented elaborations would confirm that the anomalous variability
142 of the diffusion and virulence of SARS-CoV-2 in Italy could partly depend on the levels of
143 atmospheric contamination. Although the capability of this coronavirus to bind particulate matters
144 remains to be established, chronic exposure to atmospheric contamination and related diseases
145 may represent a risk factor in determining the severity of COVID-19 syndrome and the high
146 incidence of fatal events (Chen et al., 2020; Conticini et al., 2020; Dutheil et al., 2020; Wang D. et
147 al., 2020; Wu et al., 2020).

148 In conclusion, the actual pandemic event is demonstrating that infectious diseases
149 represent one of the key challenges for human society. Since periodic emergence of viral agents
150 show an increasing correlation with socio-economic, environmental and ecological factors
151 (Morens et al., 2004; Jones et al., 2008), also air quality should be considered as part of an
152 integrated approach toward sustainable development, human health protection and prevention of
153 epidemic spreads. However, the role of atmospheric pollution should be considered in a long-
154 term, chronic perspective, and adoption of mitigation actions only during a viral outbreak could be
155 of limited utility.

156 **Acknowledgement.** No funding was received for this study.

157

158 **Note:** *A complete description of the origin of the used data and the methods of graphic and*
159 *statistical processing is included in the Supplementary Materials.*

References

1. Brunekreef, B., & Holgate, S. T. (2002). Air pollution and health. *The Lancet*, 360(9341), 1233–1242. [https://doi.org/10.1016/s0140-6736\(02\)11274-8](https://doi.org/10.1016/s0140-6736(02)11274-8)
2. Cereda, D., Tirani, M., Rovida, F., Demicheli, V., Ajelli, M., Poletti, P., Trentin, F., Guzzetta, G., Marziano, V., Barone, A., Magoni, M., Deandrea, S., Diurno, G., Lombardo, M., Faccini, M., Pan, A., Bruno, R., Pariani, E., Grasselli, G., Piatti, A., Gramegna, M., Baldanti, F., Melegaro, A., Merler, S. (2020). The early phase of the COVID-19 outbreak in Lombardy, Italy. *Arxiv pre-print*. <https://arxiv.org/ftp/arxiv/papers/2003/2003.09320.pdf>
3. Chen, G., Zhang, W., Li, S., Williams, G., Liu, C., Morgan, G. G., Jaakkola, J. J. K., & Guo, Y. (2017). Is short-term exposure to ambient fine particles associated with measles incidence in China? A multi-city study. *Environmental Research*, 156, 306–311. <https://doi.org/10.1016/j.envres.2017.03.046>
4. Chen, N., Zhou, M., Dong, X., Qu, J., Gong, F., Han, Y., ... Zhang, L. (2020). Epidemiological and clinical characteristics of 99 cases of 2019 novel coronavirus pneumonia in Wuhan, China: a descriptive study. *The Lancet*, 395(10223), 507–513. [https://doi.org/10.1016/s0140-6736\(20\)30211-7](https://doi.org/10.1016/s0140-6736(20)30211-7)
5. Chen, P.-S., Tsai, F. T., Lin, C. K., Yang, C.-Y., Chan, C.-C., Young, C.-Y., & Lee, C.-H. (2010). Ambient Influenza and Avian Influenza Virus during Dust Storm Days and Background Days. *Environmental Health Perspectives*, 118(9), 1211–1216. <https://doi.org/10.1289/ehp.0901782>
6. Ciencewicki, J., & Jaspers, I. (2007). Air Pollution and Respiratory Viral Infection. *Inhalation Toxicology*, 19(14), 1135–1146. <https://doi.org/10.1080/08958370701665434>
7. Cohen, J., & Kupferschmidt, K. (2020). Countries test tactics in “war” against COVID-19. *Science*, 367(6484), 1287–1288. <https://doi.org/10.1126/science.367.6484.1287>
8. Conticini, E., Frediani, B., & Caro, D. (2020). Can atmospheric pollution be considered a co-factor in extremely high level of SARS-CoV-2 lethality in Northern Italy? *Environmental Pollution*, 114465. <https://doi.org/10.1016/j.envpol.2020.114465>
9. Cucinotta, D., Vanelli, M. (2020). WHO Declares COVID-19 a Pandemic. *Acta Bio Medica*, 91(1), 157–160. <https://doi.org/10.23750/abm.v91i1.9397>
10. Dutheil, F., Baker, J.S., & Navel, V. (2020). COVID-19 as a factor influencing air pollution? *Environmental Pollution*, 263, 114466. <https://doi.org/10.1016/j.envpol.2020.114466>
11. ECDC, European Centre for Disease Prevention and Control (European Union Agency), Situation update worldwide, as of 7 April 2020. <https://www.ecdc.europa.eu/en/geographical-distribution-2019-ncov-cases>
12. EEA, European Environmental Agency (2019). Air quality in Europe, Report Number 10/2019.
13. EEA, European Environmental Agency (2020). Monitoring Covid-19 impacts on air pollution. Dashboard Prod-ID: DAS-217-en. <https://www.eea.europa.eu/themes/air/air-quality-and-covid19/monitoring-covid-19-impacts-on>

14. Fanelli, D., & Piazza, F. (2020). Analysis and forecast of COVID-19 spreading in China, Italy and France. *Chaos, Solitons & Fractals*, 134, 109761. <https://doi.org/10.1016/j.chaos.2020.109761>
15. Gouda, M. M., Shaikh, S. B., & Bhandary, Y. P. (2018). Inflammatory and Fibrinolytic System in Acute Respiratory Distress Syndrome. *Lung*, 196(5), 609–616. <https://doi.org/10.1007/s00408-018-0150-6>
16. Huang, L., Zhou, L., Chen, J., Chen, K., Liu, Y., Chen, X., & Tang, F. (2016). Acute effects of air pollution on influenza-like illness in Nanjing, China: A population-based study. *Chemosphere*, 147, 180–187. <https://doi.org/10.1016/j.chemosphere.2015.12.082>
17. ICPD, Italian Civil Protection Department (2020). COVID-19 Monitoring, daily update (accessed 8 April 2020). <https://github.com/pcm-dpc/COVID-19>
18. Jones, K. E., Patel, N. G., Levy, M. A., Storeygard, A., Balk, D., Gittleman, J. L., & Daszak, P. (2008). Global trends in emerging infectious diseases. *Nature*, 451(7181), 990–993. <https://doi.org/10.1038/nature06536>
19. Longhin, E., Holme, J. A., Gualtieri, M., Camatini, M., & Øvrevik, J. (2018). Milan winter fine particulate matter (wPM_{2.5}) induces IL-6 and IL-8 synthesis in human bronchial BEAS-2B cells, but specifically impairs IL-8 release. *Toxicology in Vitro*, 52, 365–373. <https://doi.org/10.1016/j.tiv.2018.07.016>
20. Kampa, M., & Castanas, E. (2008). Human health effects of air pollution. *Environmental Pollution*, 151(2), 362–367. <https://doi.org/10.1016/j.envpol.2007.06.012>
21. Legambiente, Annual Dossier series on Air Quality in Italy (2018, 2019 and 2020). *Mal’Aria di città, Dossier 2018, 2019 and 2020*, referring the European Environmental Agency (EEA).
22. Li, Q. et al. Early Transmission Dynamics in Wuhan, China, of Novel Coronavirus–Infected Pneumonia. *N. Engl. J. Med.* NEJMoa2001316, <https://doi.org/10.1056/NEJMoa2001316>
23. Morens, D. M., Folkers, G. K., & Fauci, A. S. (2004). The challenge of emerging and re-emerging infectious diseases. *Nature*, 430(6996), 242–249. <https://doi.org/10.1038/nature02759>
24. Peng, L., Zhao, X., Tao, Y., Mi, S., Huang, J., & Zhang, Q. (2020). The effects of air pollution and meteorological factors on measles cases in Lanzhou, China. *Environmental Science and Pollution Research*. <https://doi.org/10.1007/s11356-020-07903-4>
25. Perret, J., Bowatte, G., Lodge, C., Knibbs, L., Gurrin, L., Kandane-Rathnayake, R., Johns, D., Lowe, A., Burgess, J., Thompson, B., Thomas, P., Wood-Baker, R., Morrison, S., Giles, G., Marks, G., Markos, J., Tang, M., Abramson, M., Walters, E., ... Dharmage, S. (2017). The Dose–Response Association between Nitrogen Dioxide Exposure and Serum Interleukin-6 Concentrations. *International Journal of Molecular Sciences*, 18(5), 1015. <https://doi.org/10.3390/ijms18051015>

26. Remuzzi, A., & Remuzzi, G. (2020). COVID-19 and Italy: what next? *The Lancet*, 395(10231), 1225–1228. [https://doi.org/10.1016/s0140-6736\(20\)30627-9](https://doi.org/10.1016/s0140-6736(20)30627-9)
27. Sedlmaier, N., Hoppenheidt, K., Krist, H., Lehmann, S., Lang, H., & Büttner, M. (2009). Generation of avian influenza virus (AIV) contaminated fecal fine particulate matter (PM2.5): Genome and infectivity detection and calculation of immission. *Veterinary Microbiology*, 139(1–2), 156–164. <https://doi.org/10.1016/j.vetmic.2009.05.005>
28. Shinya, K., Ebina, M., Yamada, S., Ono, M., Kasai, N., & Kawaoka, Y. (2006). Influenza virus receptors in the human airway. *Nature*, 440(7083), 435–436. <https://doi.org/10.1038/440435a>
29. Sohrabi, C., Alsafi, Z., O'Neill, N., Khan, M., Kerwan, A., Al-Jabir, A., Iosifidis, C., & Agha, R. (2020). World Health Organization declares global emergency: A review of the 2019 novel coronavirus (COVID-19). *International Journal of Surgery*, 76, 71–76. <https://doi.org/10.1016/j.ijisu.2020.02.034>
30. van Riel, D., Munster, V. J., de Wit, E., Rimmelzwaan, G. F., Fouchier, R. A. M., Osterhaus, A. D. M. E., & Kuiken, T. (2006). H5N1 Virus Attachment to Lower Respiratory Tract. *Science*, 312(5772), 399–399. <https://doi.org/10.1126/science.1125548>
31. Wang, D., Hu, B., Hu, C., Zhu, F., Liu, X., Zhang, J., ... Peng, Z. (2020). Clinical Characteristics of 138 Hospitalized Patients With 2019 Novel Coronavirus–Infected Pneumonia in Wuhan, China. *JAMA*, 323(11), 1061. <https://doi.org/10.1001/jama.2020.1585>
32. Wang, T., Du, Z., Zhu, F., Cao, Z., An, Y., Gao, Y., & Jiang, B. (2020). Comorbidities and multi-organ injuries in the treatment of COVID-19. *The Lancet*, 395(10228), e52. [https://doi.org/10.1016/s0140-6736\(20\)30558-4](https://doi.org/10.1016/s0140-6736(20)30558-4)
33. Watts, N., Amann, M., Arnell, N., Ayeb-Karlsson, S., Belesova, K., Boykoff, M., Byass, P., Cai, W., Campbell-Lendrum, D., Capstick, S., Chambers, J., Dalin, C., Daly, M., Dasandi, N., Davies, M., Drummond, P., Dubrow, R., Ebi, K. L., Eckelman, M., ... Montgomery, H. (2019). The 2019 report of The Lancet Countdown on health and climate change: ensuring that the health of a child born today is not defined by a changing climate. *The Lancet*, 394(10211), 1836–1878. [https://doi.org/10.1016/s0140-6736\(19\)32596-6](https://doi.org/10.1016/s0140-6736(19)32596-6)
34. WHO Director, General's Remarks at the Media Briefing on 2019-nCoV on 11 February 2020 (2020a). <https://www.who.int/dg/speeches/detail/who-director-general-s-remarks-at-the-media-briefing-on-2019-ncov-on-11-february-2020>.
35. WHO Director, General's opening remarks at the media briefing on COVID19 on 11 March 2020 (2020b). <https://www.who.int/dg/speeches/detail/who-director-general-s-opening-remarks-at-the-media-briefing-on-covid-19---11-march-2020>
36. Wu, F., Zhao, S., Yu, B., Chen, Y.-M., Wang, W., Song, Z.-G., ... Zhang, Y.-Z. (2020). A new coronavirus associated with human respiratory disease in China. *Nature*, 579(7798), 265–269. <https://doi.org/10.1038/s41586-020-2008-3>

37. Xu, B., Gutierrez, B., Mekaru, S., Sewalk, K., Goodwin, L., Loskill, A., Cohn, E. L., Hswen, Y., Hill, S. C., Cobo, M. M., Zarebski, A. E., Li, S., Wu, C.-H., Hullah, E., Morgan, J. D., Wang, L., O'Brien, K., Scarpino, S. V., Brownstein, J. S., ... Kraemer, M. U. G. (2020). Epidemiological data from the COVID-19 outbreak, real-time case information. *Scientific Data*, 7(1). <https://doi.org/10.1038/s41597-020-0448-0>
38. Ye, Q., Fu, J., Mao, J., & Shang, S. (2016). Haze is a risk factor contributing to the rapid spread of respiratory syncytial virus in children. *Environmental Science and Pollution Research*, 23(20), 20178–20185. <https://doi.org/10.1007/s11356-016-7228-6>
39. Zhu, Y., Xie, J., Huang, F., & Cao, L. (2020). Association between short-term exposure to air pollution and COVID-19 infection: Evidence from China. *Science of the Total Environment*, 727,138704. <https://doi.org/10.1016/j.scitotenv.2020.138704>

Legend of the Figures:

Figure 1 - Regional distribution of COVID-19 outbreak in Italy (from 24 February to 27 April 2020. A) abundance of COVID-19 cases (absolute number); B) percentage of positive subjects referred to the number of performed tests (oropharyngeal swabs); C) mortality rate on the number of positive cases; D) percentage of COVID-19 cases normalized to the number of inhabitants; E) percentage of deaths normalized to the number of inhabitants. Data obtained and re-elaborated from the official daily reports of Italian Civil Protection Department (ICPD, 2020).

Figure 2 - Regional data on air quality levels: A) nitrogen dioxide (NO₂); B) particulate matter of 2.5 µm or less (PM2.5); C) particulate matter of 2.5-10 µm (PM10). Data are referred to the means of values of last four years (2016-2019), expressed as µg/m³ (obtained and elaborated from the European Environmental Agency, accessed on 6 April, EEA, 2020).

Figure 3 - Number of days per year exceeding the regulatory limits relating to A) ozone (O₃) and to B) particulate matter (PM10), as average means of the last 3 years (2017-2019); C) number of years in which the PM10 limit was exceeded for at least 35 days per year, from 2010 to 2019. Data are obtained and elaborated from annual reports (Legambiente, 2018; 2019; 2020) and referred to the official statistics of the European Environmental Agency (EEA, 2019).

Figure 4 - Statistical correlation between the regional distribution of COVID-19 cases and the air quality parameters in Italy: incidence of COVID-19 cases vs levels of A) NO₂, B) PM2.5 and C) PM10 (four years means); incidence of COVID-19 cases vs number of days exceeding regulatory limits of D) O₃ and E) PM10 (three years means). Data obtained and elaborated from EEA (2019; 2020), ICPD (2020), and Legambiente (2018; 2019; 2020).

Figure 1.

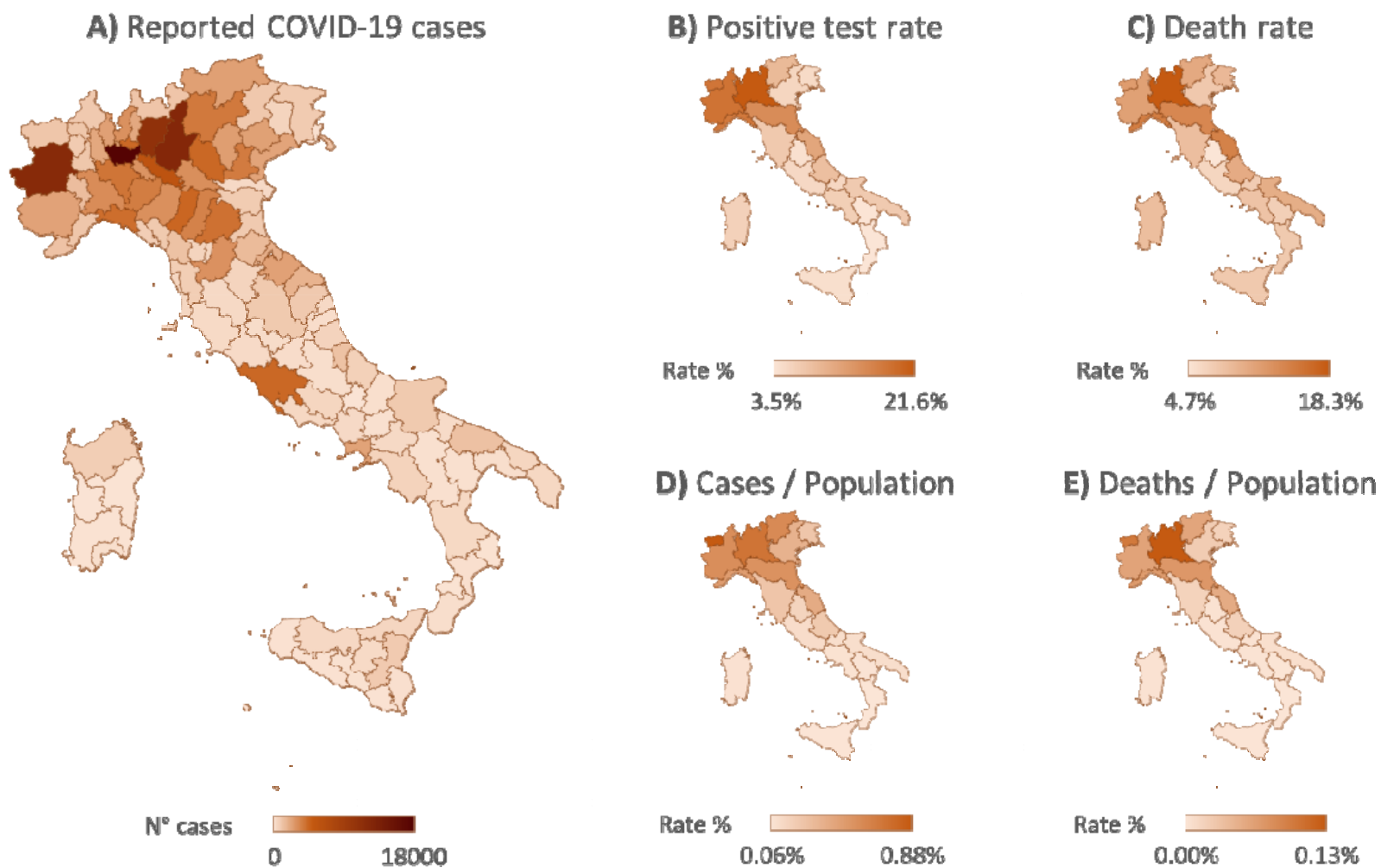


Figure 2.

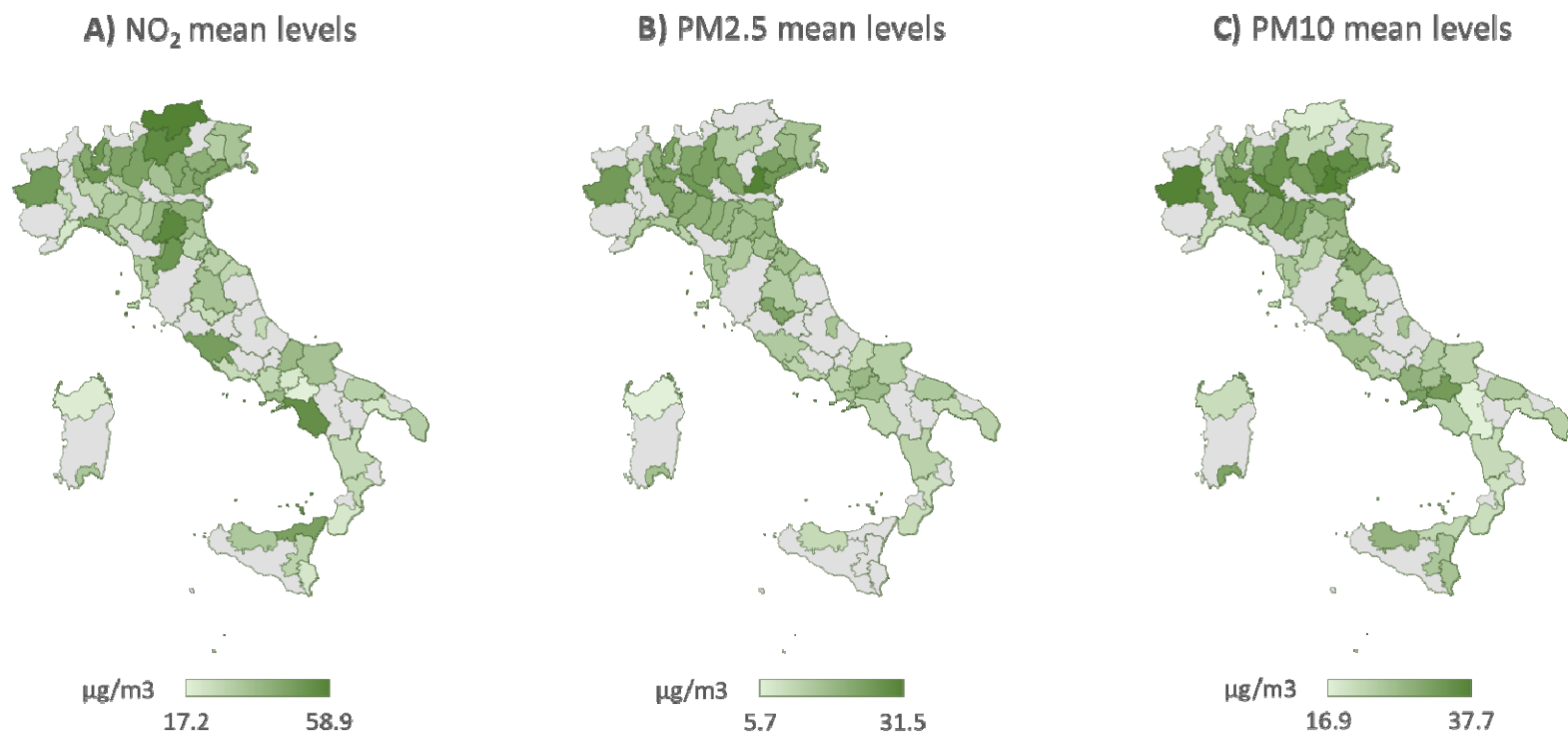


Figure 3.

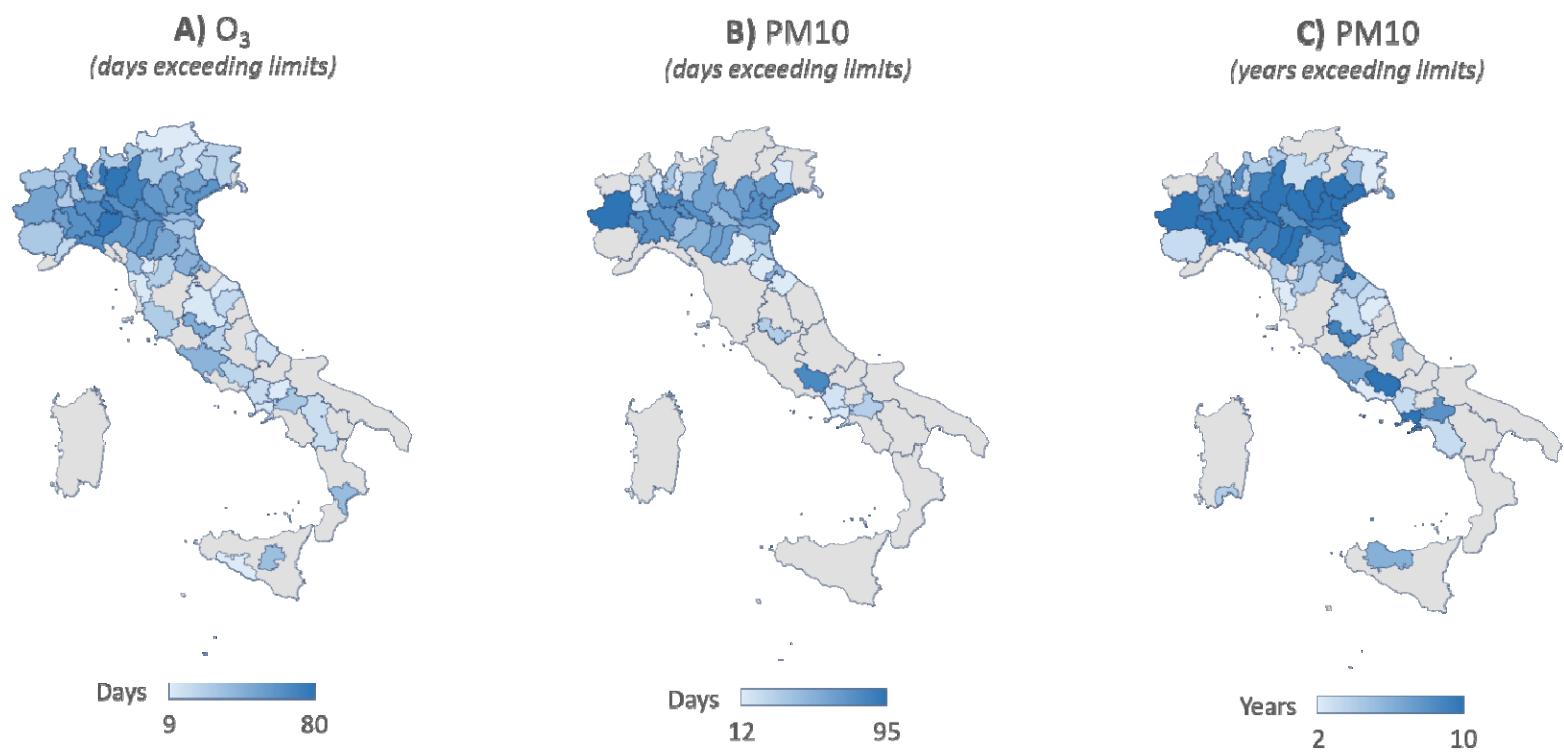
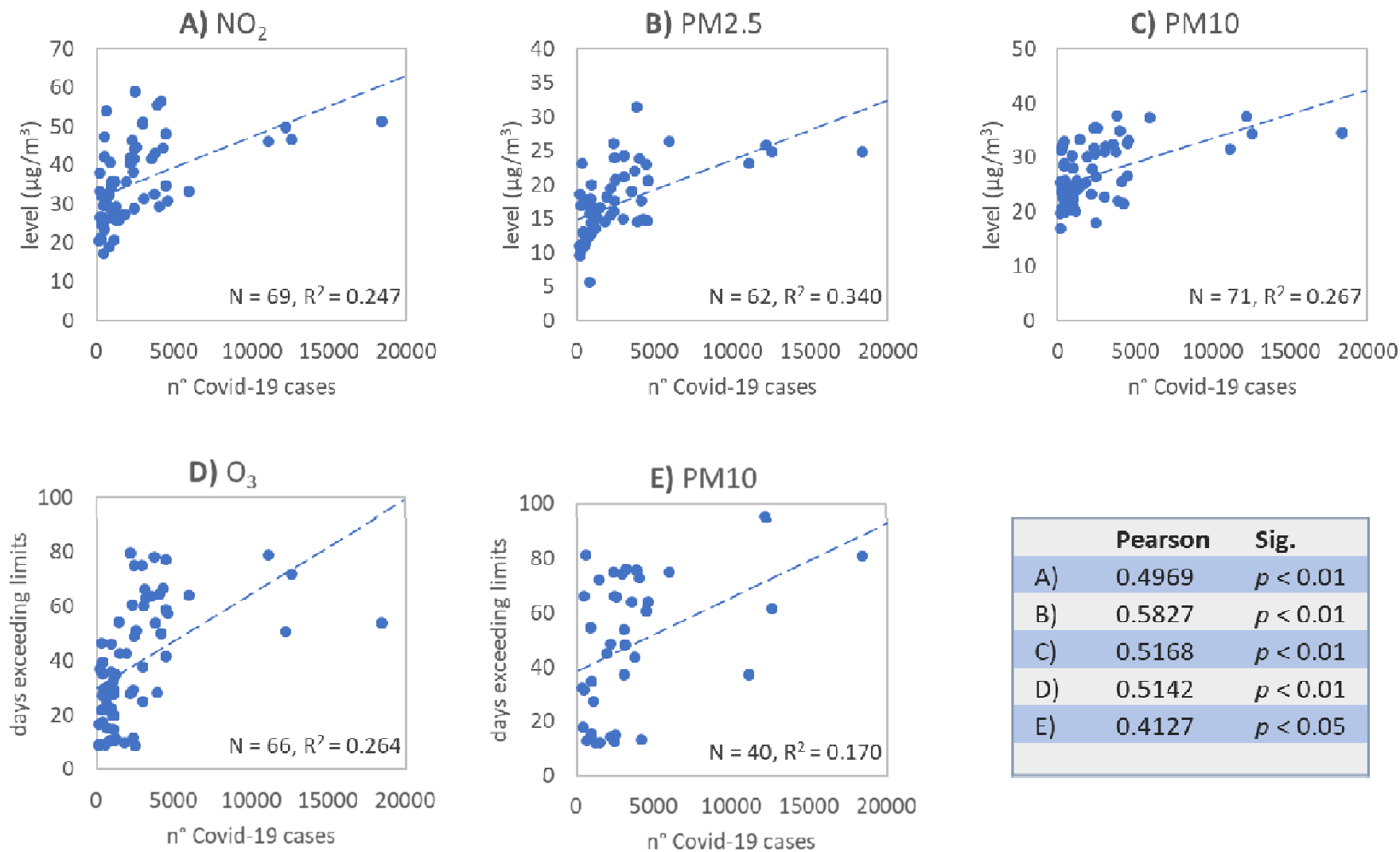
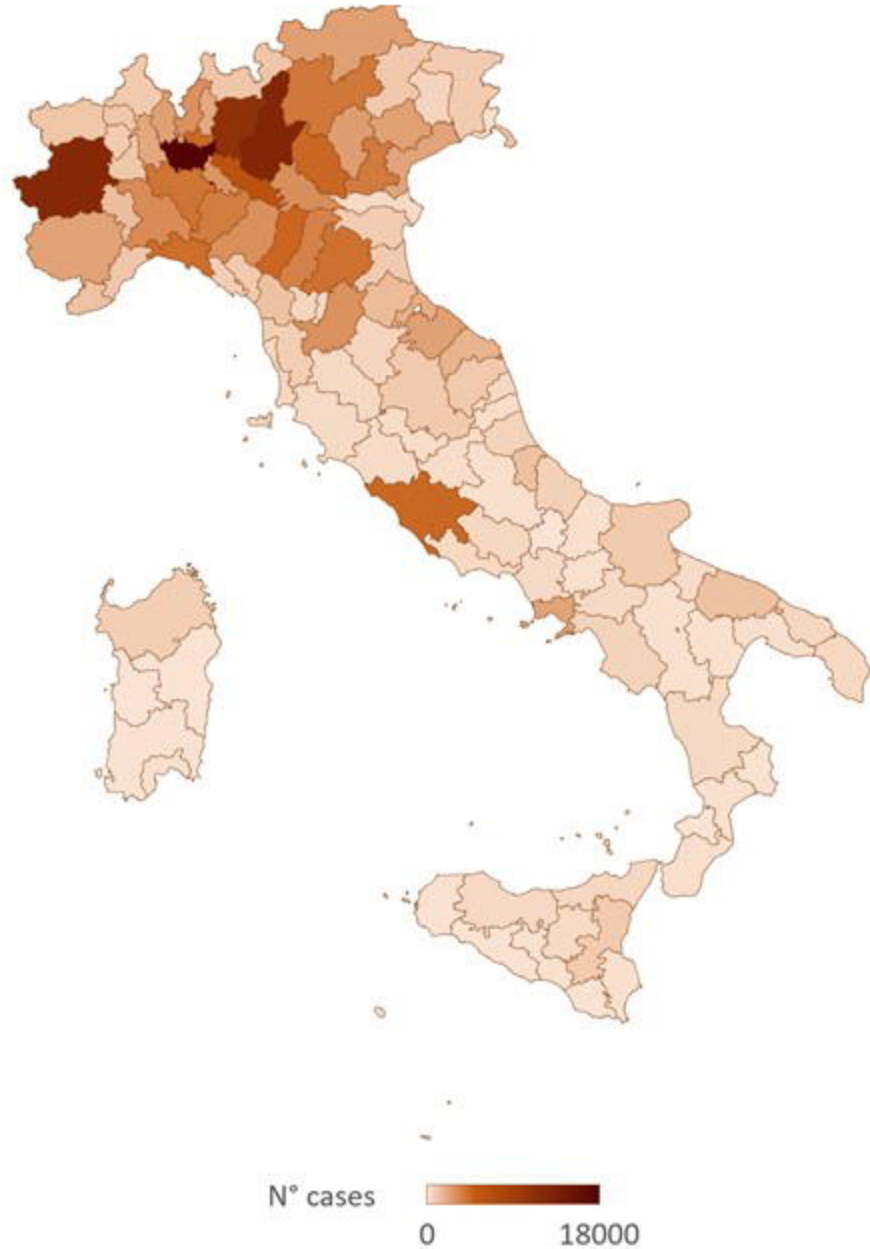


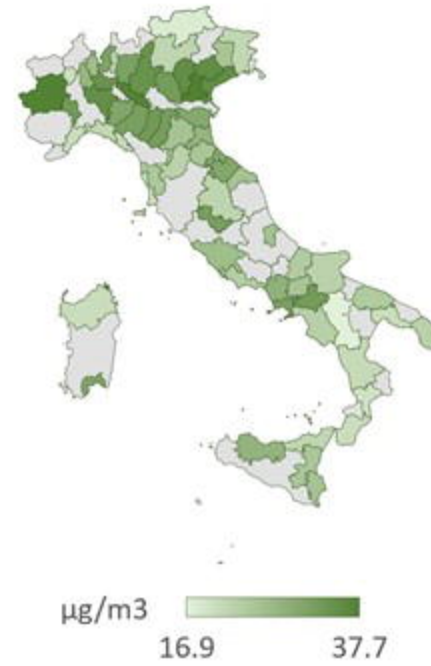
Figure 4.



Reported cases of COVID-19
(updated April 27)



Mean levels of PM₁₀
in the last 4 years



Days/year exceeding limits of
PM₁₀ (last 3 years)

