

1 **Title:** Adsorption of respiratory syncytial virus (RSV), rhinovirus, SARS-CoV-2, and F+
2 bacteriophage MS2 RNA onto wastewater solids from raw wastewater

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11

12 **Abstract**

13 Despite the wide adoption of wastewater surveillance, more research is needed to understand
14 the fate and transport of viral genetic markers in wastewater. This information is essential for the
15 interpretation of wastewater surveillance data and the development of mechanistic models that
16 link wastewater measurements to the number of individuals shedding virus. In this study, we
17 examined the solid-liquid partitioning behavior of four viruses in wastewater: SARS-CoV-2,
18 respiratory syncytial virus (RSV), rhinovirus (RV), and F+ coliphage/MS2. We used two
19 approaches to achieve this: we (1) conducted laboratory partitioning experiments using lab-
20 grown viruses and (2) examined the distribution of endogenous viruses in wastewater. Partition
21 experiments were conducted at 4°C and 22°C; wastewater samples were spiked with varying
22 concentrations of each virus and stored for three hours to allow the system to equilibrate. Solids
23 and liquids were separated via centrifugation and viral RNA concentrations were quantified
24 using reverse-transcription-digital droplet PCR (RT-ddPCR). For the distribution experiment,
25 wastewater samples were collected from six wastewater treatment plants and processed
26 without spiking exogenous viruses; viral RNA concentrations were measured in wastewater

27 solids and liquid. Overall, RNA concentrations were higher in solids than the liquid fraction of
28 wastewater by approximately 3–4 orders of magnitude. Partition coefficients (K_F) from laboratory
29 experiments were determined using the Freundlich model and ranged from 2,000–270,000 ml·g⁻¹
30 across viruses and temperature conditions. Distribution coefficients (K_d) determined from
31 endogenous wastewater viruses were consistent with results from laboratory experiments.
32 Further research is needed to understand how virus and wastewater characteristics might
33 influence the partition of viral genetic markers in wastewater.

34

35 **Keywords:** virus, partitioning, wastewater, RSV, rhinovirus, SARS-CoV-2, MS2, F+ coliphage

36

37 **Synopsis:** We examined the solid-liquid partitioning behavior of SARS-CoV-2, RSV, RV, and
38 F+coliphage/MS2 RNA in wastewater influent. Overall, partition/distribution coefficients were
39 similar across viruses and temperature conditions.

40

41 **Introduction**

42 Multiple countries are currently monitoring the spread of COVID-19 by measuring the genetic
43 markers of SARS-CoV-2 variants in wastewater and primary settled solids (hereafter referred to
44 as wastewater matrices). A few wastewater surveillance programs also monitor the genetic
45 markers of common respiratory diseases like influenza, respiratory syncytial virus (RSV),
46 rhinovirus (RV), and human metapneumovirus (HMPV).^{1–5} This information can be used by
47 public health officials to monitor infection trends, complement clinical surveillance data, and
48 strengthen public health responses.⁶ Despite the widespread adoption of wastewater
49 surveillance, research is still needed to understand the fate and transport of viral genetic
50 markers in wastewater matrices. This information is essential for linking wastewater surveillance
51 data to the number of individuals in sewersheds shedding virus, and disease incidence and
52 prevalence.^{7–9}

53

54 Viral adsorption can be influenced by the physical, chemical, and biological characteristics of
55 wastewater (e.g., temperature, pH, organic matter) and characteristics of viruses (e.g., virus
56 structure and size).^{10–13} Previous studies suggest that viruses and their genetic markers tend to
57 partition more favorably to the solid fraction of wastewater matrices than the liquid fraction.^{2,14–17}
58 For example, Mercier et al.² studied the distribution of endogenous influenza A virus (IAV) in
59 wastewater influent and primary sludge and found that the majority of IAV RNA was in settled
60 solids compared to suspended solids (larger than 0.45 μm) and the liquid fraction of these
61 matrices. Li et al.¹⁶ also examined the distribution of endogenous SARS-CoV-2 RNA (N1, N2,
62 and E gene targets) in wastewater influent and found that the majority of viral genetic markers
63 were in the solid fraction of wastewater. A few studies have also reported higher concentrations
64 of viral genetic markers in primary sludge compared to paired wastewater influent samples. For
65 instance, Wolfe et al.¹⁸ found that IAV RNA concentrations were 1,000 times higher in primary
66 settled solids than wastewater influent on mass equivalent basis. Similar results have been
67 reported for SARS-CoV-2, MPOX virus, and pepper mild mottle virus (PMMoV) RNA where viral
68 RNA concentrations were enriched by 3-4 orders of magnitude in primary settled solids
69 compared to wastewater influent.^{14–16,19} Yin et al.²⁰ summarized the solid-liquid distribution of
70 different strains of enteroviruses, hepatitis A, adenovirus, rotavirus, and bacteriophages in
71 wastewater and activated sludge and found that viral adsorption can vary greatly between
72 viruses and wastewater matrices. Still, in all cases, viruses tended to partition to wastewater
73 solids.

74

75 A few studies have also examined the equilibrium and kinetic adsorption of viruses and their
76 genetic markers in wastewater matrices. For example, Ye et al.¹² studied the sorption kinetics of
77 four infectious lab-grown human virus surrogates (MHV, $\phi 6$, MS2, and T3) in wastewater
78 influent and found that enveloped viruses partitioned more to the solid fraction of wastewater

79 compared to non-enveloped viruses. A similar study was conducted by Yang et al.¹¹ but using
80 molecular methods (quantitative PCR) to quantify the concentrations of four lab-grown
81 surrogates (Phi6, MS2, T4, and Phix174) in activated sludge. Partition coefficients (converted
82 from log K_F ; also known as the Freundlich coefficient) were 4.1×10^6 , 5.4×10^5 , 1.2×10^5 , and 8.5
83 $\times 10^3$ mL·g⁻¹ for Phi6, MS2, T4, and Phix174 in sludge, respectively. Yin et al.²⁰ also measured
84 the sorption of human Adenovirus 40 (HAV40) in primary and secondary sludge and found that
85 the majority of HAV40 DNA was adsorbed into the solids fraction of these matrices. Partition
86 coefficients (reported as K_p in the paper) were 3.7×10^4 mL·g⁻¹ and 4.0×10^4 mL·g⁻¹ in primary and
87 secondary sludge, respectively. Researchers have also examined the equilibrium and kinetic
88 adsorption of SARS-CoV-2 RNA onto passive samplers designed for wastewater surveillance.²¹
89
90 In this study, we examined the partitioning behavior of four viruses in wastewater: SARS-CoV-2,
91 respiratory syncytial virus (RSV), rhinovirus (RV), and MS2/F+ coliphage. We achieve this
92 through laboratory partitioning experiments and through examination of the distribution of these
93 viruses in actual wastewater samples. SARS-CoV-2, RSV, and RV were chosen for the study
94 because their equilibrium partitioning and distribution in wastewater has not been previously
95 studied. MS2 was chosen because it is widely used as a surrogate for pathogenic respiratory
96 viruses in lab experiments. These viruses represent both enveloped (SARS-CoV-2 and RSV)
97 and non-enveloped (RV and MS2) viruses. Additionally, the human pathogenic viruses chosen
98 in this study are targets for wastewater-based epidemiology monitoring efforts. Understanding
99 the partitioning behavior of viral genetic markers could inform wastewater sampling strategies
100 and help optimize methods for processing wastewater and primary sludge samples. Partition
101 and distribution coefficients can also help inform complex mathematical models that aim to
102 estimate or predict the number of positive cases in communities.²²

103

104 **Materials and Methods**

105 **Overview.** We conducted two sets of experiments to examine the partition and distribution,
106 respectively, of SARS-CoV-2, RSV, RV, and F+ coliphage in wastewater influent. The
107 characteristics of these viruses are shown in Table 1. The partitioning experiment was
108 conducted using lab-grown SARS-CoV-2, RSV-A, RV-B, and MS2. For these experiments,
109 wastewater influent samples were spiked with varying concentrations of each virus and
110 incubated at two different temperatures (4°C and 22°C) to allow the system to equilibrate. After
111 incubation, influent samples were centrifuged and decanted to obtain an aliquot from the liquid
112 and solid fractions. RNA was extracted from the aliquots and quantified using reverse-
113 transcription-digital droplet PCR (RT-ddPCR). The distribution experiment examined the
114 distribution of endogenous SARS-CoV-2, RSV, RV, and F+ coliphage in actual wastewater
115 samples. Influent samples were collected from six wastewater plants and processed using the
116 previously described method, but without spiking with viral surrogates. The following sections
117 provide a detailed description of the experiments. Reporting of methods follows EMMI
118 guidelines²³ (Figure S1 provides EMMI checklist and details).

119

120 **Wastewater sample collection.** The partitioning experiment was conducted using two influent
121 samples from the Palo Alto Regional Water Quality Control Plant (PA). The plant serves
122 approximately 215,000 people and treats an annual average daily flow of 19.8 million gallons
123 per day (MGD). Total suspended solids (TSS) and pH levels range from 220–360 mg/L and
124 7.5–7.8, respectively. We collected approximately two liters of a 24-hour influent composite
125 sample on September 20, 2022 for the 4°C experiment and on October 28, 2022 for the 22°C
126 experiment. Samples were collected in 10% HCl acid-washed plastic containers, stored at the
127 respective experimental temperatures, and spiked with a mixture of lab-grown SAR-CoV-2,
128 RSV-A, RV-B, and MS2 within 24 hours of sample collection (see detailed methods below) for
129 experiments.

130 **Virus purification and spike cocktails.** Heat-inactivated SARS-CoV-2 (Isolate: USA-
131 WA1/2020; catalog no. 0810587CFHI), RSV-A (catalog no. 0810040ACF), and RV-B (catalog
132 no. 0810284CF) were purchased from ZeptoMetrix (Buffalo, New York). The manufacturer
133 inactivates SARS-CoV-2 by heating the virus at 60°C for 1 hour. RSV-A and RV-B are viable
134 viruses suspended in cell culture fluids. *Escherichia coli* phage MS2 (DMS no. 13767) was
135 purchased from the DSMZ German Collection of Microorganisms and Cell Cultures. Viruses
136 were purified to remove viral culture fluid using Amicon® Ultra-0.5 ml centrifugal filters (100 kDa
137 MWCO; Millipore UFC5100) following the manufacturer's instructions. Briefly, 0.5 ml of virus
138 stock, as received from the vendor, was added to individual centrifugal filters and centrifuged at
139 14,000xg for 5 min. The filters were immediately flipped and centrifuged at 1,000xg for 2
140 minutes to recover the retentate. A series of dilutions were prepared using autoclaved
141 phosphate-buffered saline (PBS; Fisher BioReagents, Pittsburgh, Pennsylvania) to achieve a
142 viral genome concentration of approximately 1×10^3 , 1×10^4 , 1×10^5 , 1×10^6 , and 1×10^7 cp/μl PBS.
143 A total of five spike cocktails were prepared by mixing equal volumes of purified heat-inactivated
144 SARS-CoV-2, RSV-A, RV-B, and MS2 stock. The final concentrations of the five stock cocktails
145 ranged from approximately 1×10^3 – 1×10^7 cp/μl PBS (10^3 , 10^4 , 10^5 , 10^6 , 10^7) for each virus.

146 **Preanalytical processing.** Wastewater samples were thoroughly mixed by inverting 3–4 times
147 and aliquoted into eighteen 50 ml centrifuge tubes (hereafter, referred to as subsamples). Five
148 sets of three subsamples were then spiked with one of the five different concentrations of spiked
149 cocktails to achieve a final concentration of approximately 10^3 , 10^4 , 10^5 , 10^6 , or 10^7 cp/ml for
150 each virus. Three additional subsamples were reserved to measure the background
151 concentration of endogenous SARS-CoV-2, RSV, RV, and F+ coliphage RNA. After spiking,
152 subsamples were stored at 4°C or 22°C, depending on the experiment, and gently mixed (~20
153 rpm) using a tube roller (Globe Scientific, GSCI-GTR-AVS) for approximately three hours to
154 allow the system to equilibrate. The time needed to reach equilibrium (< 3 hours) was

155 determined based on a preliminary experiment (see Figure S2 and S3) and it is consistent with
156 other virus adsorption studies conducted in wastewater.^{11,12} After three hours, subsamples were
157 centrifuged at 24,500xg for 20 minutes at the temperatures of the experiments. This process
158 removes solid particles with hydrodynamic radii greater than 0.3 μm in diameter.²⁴ 200 μl of the
159 supernatant was transferred to a 2 ml collection tube and spiked with 5 μl of bovine coronavirus
160 vaccine (BCoV; Zoetis, CALF-GUARD®; Parsippany-Troy Hills, NJ) as an extraction recovery
161 control. The BCoV vaccine comes lyophilized and was resuspended in 3 mL of molecular-grade
162 water. These aliquots represent the liquid fraction of the wastewater sample.

163 The remaining supernatant was decanted and the pellet represents dewatered solids. A portion
164 of the dewatered solids was reserved to calculate the dry weight. Dewatered solids were
165 weighed before and after drying at 105°C for 24 hours. To collect solids for viral analysis,
166 approximately 0.1 g of dewatered solids were collected from the bottom of the centrifuge tube
167 and aliquoted into 2 ml microcentrifuge tubes using a disposable spatula (Fisher Scientific,
168 catalog no. 50-476-569). Dewatered solids were resuspended in BCoV-spiked DNA/RNA shield
169 (Zymo Research, catalog no. R1100-250) at a concentration of approximately 75 mg/ml; this
170 concentration of solids in the buffer has been shown to alleviate downstream RT-ddPCR
171 inhibition.²⁵ Spiked DNA/RNA shield was pre-prepared using 1.5 μl of BCoV per ml of DNA/RNA
172 shield. Three to five grinding balls (OPS DIAGNOSTICS, GBSS 156-5000-01) were added to a
173 2 ml microcentrifuge tube and the mixture was homogenized at 4 m/s for 1 minute using the MP
174 Bio Fastprep-24TM (MP Biomedicals, Santa Ana, CA). The homogenized aliquots were then
175 centrifuged for 5 min at 5,250 x g and 200 μl of the supernatant was transferred to a 2 ml
176 collection tube. This aliquot contains viral targets from the solid fraction of the wastewater
177 sample. Liquid and solid aliquots were stored at 4°C overnight and the nucleic acids were
178 extracted from them the next day.

179 **RNA extraction.** RNA was extracted from the solid and liquid aliquots using the Qiagen AllPrep
180 PowerViral DNA/RNA kit and further purified using the Zymo OneStep PCR inhibitor removal
181 columns (Zymo Research, Irvine, CA) following the manufacturer's instructions. RNA extracts
182 were aliquoted into 1.5 ml DNA LoBind tubes, stored at -80°C for less than two weeks, and
183 thawed once (1 freeze-thaw cycle) before quantification. Nuclease-free water and BCoV-spiked
184 DNA/RNA shield were used as negative and positive extraction controls, respectively. These
185 controls were carried through the extraction process with one set of positive and negative
186 controls per extraction batch (~15 solid or liquid aliquots/batch of extraction).

187 **RNA Quantification.** SARS-CoV-2 and RSV were quantified using a duplex assay described in
188 Hughes et al.²⁶ and RV, MS2, and BCoV were quantified using singleplex assays from previous
189 studies.^{1,15,27} Primers and probes were purchased from Integrated DNA Technologies (IDT, San
190 Diego, CA) and are provided in Table S1. Using NCBI Blast, we determined that the MS2 assay
191 will detect, in addition to MS2, the following sequenced and deposited strains of genotype group
192 1 (GI) F+ RNA coliphages: JP501, M12, DL16, DL52, and DL54 which could potentially be
193 present in wastewater. There could be additional unsequenced/undeposited endogenous
194 wastewater F+ RNA coliphages that are also detectable using this assay. Therefore, detections
195 with the MS2 in wastewater will be referred to as F+ coliphage, hereafter.

196 All viral targets were quantified using the One-Step RT-ddPCR Advanced Kit for Probes (Bio-
197 Rad 1863021). For the RT-ddPCR, 20 µl of a 22 µl reaction mix was prepared for each well.
198 The reaction consisted of 5.5 µl of template RNA, 5.5 µl of Supermix, 2.2 µl of 200 U/µl Reverse
199 Transcriptase (RT), 1.1 µl of 300 mM dithiothreitol (DDT), 4.4 µl of nuclease-free water, and 3.3
200 µl of primer and probe mixture with a final concentration of 900 nM and 250 nM, respectively.
201 RNA extracts were processed undiluted and in duplicate (two technical replicates). Nuclease-
202 free water and viral RNA extracts for each target were used as negative and positive PCR
203 controls, respectively, and each run in three wells per plate. Positive controls were extracted

204 from the SARS-CoV-2, RSV, RV, and MS2 purified stocks; which are the same virus stocks
205 used to prepare the spike cocktails. Droplets were generated using the AutoDG Automated
206 Droplet Generator (Bio-Rad, Hercules, CA) and amplified using the C1000 Touch™ Thermal
207 Cycler (Bio-Rad, Hercules, CA). Thermal cycling conditions for each RT-ddPCR assay are
208 shown in Table S2. After amplification, droplets were analyzed using the QX200 droplet reader
209 and the Quantasoft Analysis Pro Software. Wells with less than 10,000 droplets were excluded
210 and technical PCR replicates (wells) were merged before performing the dimensional analysis.
211 Merged wells needed to have at least three positive droplets to be considered positive The
212 estimated lower measurement limit for solids and liquid aliquots were 3,500 cp/g and 0.7 cp/ml,
213 respectively.

214 **Distribution of endogenous viruses in wastewater.** The second experiment examined the
215 distribution of endogenous SAR-CoV-2, RSV, RV, and F+ coliphage between liquid and solid
216 fractions in wastewater samples. Influent samples (~1 L of a 24-hour composite sample) were
217 collected from six wastewater treatment plants on November 18, 2022. Samples were stored at
218 4°C and processed within 24 hours. These plants are part of an ongoing wastewater
219 surveillance program and include the following: Oceanside Water Pollution Control Plant (OS),
220 Southeast Water Pollution Control Plant (SE), Silicon Valley Clean Water Wastewater
221 Treatment Plant (SV), Sunnyvale Water Pollution Control Plant (SU), San Jose-Santa Clara
222 Regional Wastewater Facility (SJ), and South County Regional Wastewater Treatment Plant
223 (GI). The location, estimated population served, annual daily average flow, pH, and TSS for
224 each plant can be found in Table S3. Figure S4 provides a map of the sewershed (area served)
225 for each wastewater treatment plant.

226 Wastewater samples were thoroughly mixed, poured into 50 ml centrifuge tubes, and
227 processed. Three subsamples were prepared for each wastewater treatment plant, for a total of
228 18 subsamples. Solid and liquid aliquots were obtained from the wastewater samples and

229 nucleic acids were extracted. Viral targets were quantified from the aliquots using the methods
230 described above for the partitioning experiments.

231 **Dimensional analysis, adsorption models, and statistical analysis.** SARS-CoV-2, RSV, RV,
232 and MS2/F+ coliphage RNA concentrations were expressed in units of copies per gram dry
233 weight solids (cp/g) or per ml liquid (cp/ml) for the solid and liquid fractions of wastewater,
234 respectively, using dimensional analysis. BCoV recovery, used as an extraction and inhibition
235 control, was calculated for the solid and liquid fractions as follows:

$$236 \quad BCoV \text{ recovery } (\%) = \frac{\text{number of gene copies in solids OR liquid fraction}}{\text{number of gene copies spiked DNA/RNA shield OR liquid fraction}} \times 100 \quad (1)$$

237 For the partition experiment using lab-grown viruses, 36 RNA concentrations (solids: N=18 and
238 liquids: N=18) were obtained for each virus and isotherm experiment (4°C and 22°C). These
239 concentrations represent triplicate measurements for six different initial conditions: one
240 unspiked and five different spiked concentrations of viruses. For each virus and temperature
241 condition, average viral RNA concentrations (solids: N=6 and liquids: N=6) and standard
242 deviations were calculated across triplicate subsamples. To calculate the recovery of spiked
243 viruses, we first subtracted the background concentrations of viral genetic markers from liquid
244 and solid fractions of spiked subsamples and estimated the total RNA recovery in spiked
245 subsample as follows:

$$246 \quad RNA \text{ recovery } (\%) = \frac{\text{conc.in solids} \times \text{mass of solids} + \text{conc.in liquid} \times \text{volume in liquid sample}}{\text{total number of gene copies spiked into sample}} \times 100$$

247 (2)

248 Partition experiment viral RNA concentrations in the solid and liquid fractions of wastewater
249 were fit to Linear, Langmuir, and Freundlich isotherm models. These models have been used in
250 previous virus partitioning studies^{11,14,20,21} and describe a multi-layer (Freundlich) or monolayer
251 (Langmuir) adsorption process. A linear model is generally used when the coverage ratio of

252 adsorption sites is minimal.²⁸ The Linear, Freundlich, and Langmuir isotherm parameters were
253 determined for each virus and temperature condition. The average relative error (ARE) was
254 calculated for each model and compared to identify the model with the best fit. The Freundlich
255 model produced the smallest ARE overall and therefore is discussed in the main paper (see
256 Table S4 for results of other models). The nonlinear and linear forms of the Freundlich model
257 are described as follows:

$$258 \quad q_e = K_F C_e^n \quad (3)$$

$$259 \quad \log q_e = \log K_F + n \log C_e \quad (4)$$

260
261 where q_e is the equilibrium concentration of viral genomes in solids (cp/g), C_e is the equilibrium
262 concentration of viral genomes in the liquid fraction of wastewater, K_F is the Freundlich constant,
263 and n is the adsorption intensity. A higher n indicates a stronger interaction between the
264 adsorbent (i.e., wastewater solids) and the adsorbate (i.e., spiked viruses). We determined the
265 Freundlich isotherm parameters, $\log K_F$ (y-intercept) and n (slope), using a linear regression
266 model (lm function) in R. Standard errors (SE) were also obtained from the linear regression
267 model in R (summary function). Finally, we compared K_F obtained for the viruses under different
268 temperature conditions by examining whether their values and standard deviations overlapped.

269
270 For the distribution experiment (examining endogenous viruses in wastewater samples), six
271 RNA concentrations (solids: N=3 and liquids: N=3) were obtained for each virus and wastewater
272 treatment plant. Only a few liquid fractions resulted in non-detects (ND). NDs were substituted
273 with half of the lower measurement limit (0.35 cp/ml for concentrations measured in liquid
274 fractions). The average concentration and standard deviation for each virus were calculated
275 using data from the three replicate subsamples. The distribution coefficient was calculated as
276 the ratio of average viral genome concentrations detected in the solid and liquid fractions of

277 wastewater influent; $K_d=C_s/C_w$; errors were determined by propagating errors on the numerator
278 and denominator. We tested the null hypothesis that K_d were the same for endogenous viruses
279 using a Kruskal-Wallis test; $p<0.05$ was used to assess statistical significance. Statistical
280 analysis was performed in R (version 4.1.2). Finally, we compared the partition coefficients (K_F)
281 for spiked viruses at 4°C and 22°C to the distribution coefficients (K_d) for endogenous viruses in
282 wastewater for each virus to determine if K_F and K_d values and errors overlapped. We also
283 summarized the results from previous experiments measuring the concentration of viral genetic
284 markers in wastewater matrices to see if our results were consistent with previously reported K_F
285 and K_d values (see Table 4).

286

287 **Results**

288 **Extraction and PCR controls**

289 Positive and negative extraction and PCR controls were positive and negative, respectively.
290 BCoV was used as a process recovery and gross inhibition control. BCoV recoveries were
291 similar between the solid and liquid fractions of wastewater. Solids and liquid aliquots had a
292 median BCoV recovery of 0.40 and 0.35, respectively. Viral genome concentrations were not
293 adjusted by recovery given the complexities associated with estimating recovery using
294 surrogate viruses²⁹ and given that the recoveries were similar. The total recovery of spiked
295 SARS-CoV-2, RSV-A, RV-B, and MS2 RNA was approximately 35%, 45%, 22%, and 33%,
296 respectively. These recoveries are similar to those of BCoV. Although inhibition can vary
297 between assays, our previous work has determined there is minimal inhibition with the pre-
298 analytical and analytical workflow,²⁵ and the similar and high detection of BCoV and spiked
299 viruses indicates inhibition is minimal.

300

301 **Partitioning of lab-grown viruses in wastewater influent.** Endogenous SARS-CoV-2, RSV,
302 RV, and MS2/F+ coliphage RNA were detected in the liquid and solid fractions of wastewater

303 influent samples from PA (Table S5). Concentrations of endogenous virus make up 22%–95%
304 of viral RNA in subsamples spiked with the lowest concentration of virus cocktail, but otherwise
305 represent a negligible percentage of the RNA in subsamples spiked with a higher concentration
306 of virus cocktail.

307

308 At equilibrium, the results from the partitioning experiments showed that RNA concentrations of
309 spiked viruses were higher in solids than the liquid fraction of wastewater influent on a mass
310 equivalent basis, by approximately 3–4 orders of magnitude. In the 4°C and 22°C isotherm
311 experiments, q_e ranged from 1.3×10^4 – 5.6×10^7 cp/g for SARS-CoV-2, 4.2×10^3 – 1.9×10^8 cp/g for
312 RSV, 2.3×10^4 – 1.3×10^7 cp/g for RV, and 4.1×10^3 – 3.2×10^7 cp/g for MS2. In the liquid fraction, C_e
313 concentrations ranged from 0.1 – 1.3×10^3 cp/ml for SARS-CoV-2, 0.2 – 4.5×10^2 cp/ml for RSV,
314 1.6 – 4.8×10^4 cp/ml for RV, and 0.8 – 4.3×10^4 cp/ml for MS2. The reported ranges represent
315 measurements across spiking conditions and the two experimental temperatures.

316

317 Viral RNA concentrations (q_e and C_e , see Figure 1) were fit to a Linear, Freundlich, and
318 Langmuir model. The Freundlich isotherm models produced the lowest ARE compared to the
319 other models based on the calculated partition coefficient (see Table S4 for results of Linear and
320 Langmuir models). Table 3 shows the Freundlich isotherm parameters (K_F and n) for each virus
321 and temperature condition. In the 4°C experiment, K_F and n ranged from 1.8×10^3 – 3.2×10^3 ml·g⁻¹
322 and 0.66–1.24, respectively. Similar results were obtained in the 22°C experiments except for
323 the partition coefficient of SARS-CoV-2. In the 22°C experiment, K_F and n ranged from 2.0×10^3 –
324 2.7×10^5 ml·g⁻¹ and 0.64–1.32, respectively. The partition coefficient of SARS-CoV-2 in the 22°C
325 experiment was significantly higher (approximately one order of magnitude) compared to other
326 viruses and temperature conditions. However, partition coefficients were not different across
327 other viruses and temperatures (see Figure S5).

328

329

330 **Distribution of endogenous viruses in wastewater influent.** In the wastewater of six
331 wastewater treatment plants, we observed results similar to those obtained in the laboratory
332 partitioning experiment. Viral RNA concentrations were higher in solids than the liquid fraction of
333 wastewater, by approximately 3–4 orders of magnitude (see Figure 1). Across six wastewater
334 treatment plants, C_s ranged from 1.6×10^3 – 1.5×10^3 cp/g (median = 3.2×10^3 cp/g) for SARS-CoV-
335 2, 9.1×10^2 – 3.7×10^3 cp/g (median = 1.7×10^3 cp/g) for RSV, 2.1×10^3 – 1.4×10^4 cp/g (median =
336 6.0×10^3 cp/g) for RV, and 4.9×10^2 – 6.1×10^3 cp/g (median = 1.1×10^3 cp/g) for F+ coliphage. C_L
337 ranged from 0.4–1.1 cp/ml (median = 0.4 cp/ml) for SARS-CoV-2, 0.4–2.7 cp/ml (median = 0.8
338 cp/ml) for RV, and 0.4–1.3 cp/ml (median = 0.7 cp/ml) for F+ coliphage. For RSV, C_L were ND
339 across wastewater treatment plants; NDs were replaced with half of the lower measurable limit
340 (0.35 cp/ml for viral concentrations in liquid fractions) to calculate the distribution coefficient (K_d).

341

342 K_d was calculated as the ratio of C_s/C_L . Across wastewater treatment plants, K_d ranged from
343 3.5×10^3 – 1.2×10^4 ml·g⁻¹ (median = 5.0×10^3 ml·g⁻¹) for SARS-CoV-2, 5.0×10^2 – 2.0×10^3 ml·g⁻¹
344 (median = 1.3×10^3 ml·g⁻¹) for RSV, 2.1×10^3 – 1.6×10^4 ml·g⁻¹ (median = 7.6×10^3 ml·g⁻¹) for RV, and
345 4.9×10^2 – 7.4×10^3 ml·g⁻¹ (median = 1.2×10^3 ml·g⁻¹) for F+ coliphage (see Table 4). Overall, RV
346 had the largest solid-liquid distribution, followed by SARS-CoV-2, F+ coliphage, and RSV. Note
347 that K_d for RSV could be higher, but we were only able to estimate a lower bound for K_d since
348 the measurement in liquid was ND. RV K_d was statistically higher from RSV and F+ coliphage
349 K_d (Kruskal-Wallis and Dunn's post hoc test, both $p < 0.05$).

350

351

352

353 **Discussion**

354

355 This is the first batch experiment examining the solid-liquid partitioning of SARS-CoV-2, RSV,
356 and RV in wastewater and the first experiment to examine the distribution of endogenous RSV
357 and RV in this matrix. Overall, higher concentrations of viral RNA were observed in solids
358 compared to the liquid fraction of wastewater, for all viruses and temperature conditions; viral
359 RNA concentrations were higher in solids by 3–4 orders of magnitude on a mass equivalent
360 basis. Partition and distribution coefficients were also similar across viruses and temperature
361 conditions, with K_F and K_d ranging from $490 \text{ ml}\cdot\text{g}^{-1}$ – $270,000 \text{ ml}\cdot\text{g}^{-1}$. Our results are consistent
362 with previously reported partition/distribution coefficients for viral genetic markers in wastewater.
363 For example, Li et al.¹⁶ measured the distribution of endogenous SARS-CoV-2 RNA (N1, N2,
364 and E genes) in wastewater samples and found higher concentrations of SARS-CoV-2 RNA in
365 solids compared to the liquid fraction; K_d (reported in their paper as the solid-liquid concentration
366 ratio) ranged from $4,000$ – $20,000 \text{ ml}\cdot\text{g}^{-1}$. Kim et al.³⁰ also studied the distribution of endogenous
367 SARS-CoV-2 RNA (S and N genes) in wastewater samples from two K-12 schools and
368 observed similar results; viral RNA concentrations were higher in solids by three orders of
369 magnitude and K_d (reported in their paper as the concentration ratio in solid to liquid samples)
370 were $8,600 \text{ ml}\cdot\text{g}^{-1}$ and $16,000 \text{ ml}\cdot\text{g}^{-1}$ for SARS-CoV-2 N and S genes, respectively. We observed
371 similar partitioning behavior for SARS-CoV-2, RSV, RV, and MS2 RNA in our study; except for
372 the partition of SARS-CoV-2 RNA at 22°C , which was higher (by approximately one order of
373 magnitude) compared to other viruses and temperature conditions.

374

375 There are very limited studies in the literature on the effects of temperature on virus adsorption
376 to particles. One study examined how temperature may influence the isotherm and kinetic
377 adsorption process of viral genetic markers in wastewater solids: Yang et al.¹¹ examined the
378 kinetic adsorption of Phi6, MS2, T4, and Phix174 RNA in primary sludge at two temperature
379 conditions (4°C and 25°C). They found that the rate of virus adsorption increased with
380 increasing temperature (i.e., the time needed to reach equilibrium was reduced). Other studies

381 have examined how temperature affects viral adsorption to clays. Syngouna et al. examined the
382 isotherm and kinetic adsorption of infectious MS2 and Φ X174 in clay particles at 4°C and 25°C,
383 and found results similar to Yang et al.³¹ However, Bellou et al.³² measured the adsorption of
384 MS2, Φ X174, and hAdV RNA on clay particles at 4°C and 25°C and found that adsorption
385 increased with decreasing temperature for hAdV, but decreased with decreasing temperature
386 for MS2 and Φ X174. In our isotherm (equilibrium) experiments, wastewater temperature did not
387 seem to have an impact on the adsorption of viral genetic markers, except for the case of
388 SARS-CoV-2. For SARS-CoV-2, a higher partition coefficient was observed at 22°C than at 4°C.
389 The equivocal results described here suggest additional work is needed to better understand
390 how temperature affects viral adsorption.

391
392 Limited previous research suggests that the presence of a lipid envelope outside the viral
393 protein capsid of a virus may impact the solid-liquid partitioning of viruses and viral genetic
394 markers in wastewater. For example, Ye et al.¹² measured the adsorption of infectious MHV, ϕ 6,
395 MS2, and T3 in wastewater samples and found that enveloped viruses (MHV and ϕ 6) were
396 more strongly associated with solids than nonenveloped viruses (MS2 and T3). Similar results
397 were reported in the study by Yang et al.,¹¹ where they examined the solid-liquid partitioning
398 behavior of Phi6, MS2, T4, and PhiX174 RNA in primary sludge samples; researchers found that
399 the majority of viral genetic markers were in the solid fraction of primary sludge. K_F ranged from
400 8,500–4,100,000 ml·g⁻¹ for Phi6, MS2, T4, and PhiX174 RNA. In our study, we observed that the
401 partition and distribution behavior across enveloped (SARS-CoV-2 and RSV) and nonenveloped
402 (RV and F+coliphage/MS2) viruses were the same. Our results were also similar across the
403 partition and distribution experiments, which suggest that lab-grown viruses and endogenous
404 viruses for SARS-CoV-2, RSV, RV, and F+coliphage/MS2 RNA may exhibit similar solid-liquid
405 partitioning behavior in wastewater.

406

407 A few mechanistic-, statistical- and epidemiological-based models have been recently proposed
408 to estimate the number of infected individuals within a sewershed. For example, Soller et al.³³
409 developed a mechanistic model that estimates the fraction of a sewershed population actively
410 infected with SARS-CoV-2. Wolfe et al.³⁴ also developed a mass balance model that links
411 SARS-CoV-2 RNA concentrations in solids to the number of individuals shedding SAR-CoV-2
412 RNA in stool within the sewershed. These models can potentially be applied to other viruses of
413 interest however, a key parameter of these mechanistic models is the solid-liquid partitioning
414 coefficient of viral genetic markers in wastewater solids. Limited information is available on viral-
415 specific partitioning data, limiting the application of these models and the interpretation of
416 wastewater surveillance data. Our results presented here fill a knowledge gap by providing
417 information on virus partitioning that can be used in these modeling applications.

418
419 Understanding the fate and transport of viral genetic markers could inform wastewater sampling
420 strategies and help optimize methods for processing wastewater and primary sludge samples.
421 For example, a study by Kim et al¹⁴ showed that methods for processing influent and settled
422 solids have comparable sensitivity. However, settled solids might be a more advantageous
423 medium in sewersheds that have a low level of active infections because it requires less sample
424 volume compared to influent methods. Our results also suggest that viral genetic markers might
425 have similar partition/distribution coefficients across wastewater treatment plants. The results
426 from our experiments might be applicable to other wastewater treatment plants that do not have
427 local partition data available. Further research should be conducted to examine the solid-liquid
428 partitioning of other viruses of interest for WBE efforts and evaluate how solid characteristics
429 (e.g. particle size and biological/ chemical composition) might influence their partitioning
430 behavior in wastewater.

431

432 **Environmental Implications.** This study fills an important knowledge gap on the partitioning of
433 respiratory viruses in wastewater and indicates that they partition preferentially to the solids.
434 This information is useful in wastewater-based epidemiology applications (both sampling and
435 modeling) but is also useful for informing wastewater treatment unit processes. These findings
436 add to a growing body of evidence that the solids in wastewater (defined as material generally
437 larger than 0.3 μm in hydrodynamic diameter) is enriched with infectious disease targets like
438 viral RNA relative to the liquid phase on a mass-equivalent basis. (Note that intact bacteria and
439 fungi, given their sizes, automatically fall into this size class.) There has been some confusion
440 among researchers and practitioners in interpreting this statement. A volume of raw wastewater
441 contains a small mass of solids (typically on the order of 10^2 mg/L), so it can be true that most of
442 the infectious disease target (mass or number) in that volume is in the liquid phase, and that the
443 concentration of the infectious disease target is enriched orders of magnitude in the solid relative
444 to the liquid phase. The fact that infectious disease targets are enriched in the solid phase
445 indicates that sample efforts that enrich for solids and choose the solids as a measurement
446 matrix will improve the sensitivity of measurement approaches.

447

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449 Muwekma Ohlone people. We pay our respects to them and their Elders, past and present, and
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451

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597 **Table 1: Characteristics of SARS-CoV-2, RSV, RV, and MS2³⁵**

Virus	Family/Genus	Genome Type	Structure	Shape	Genome Size (kb)	Virion Size (nm)
SARS-CoV-2	Coronaviridae	+ ssRNA	enveloped	spherical	30	50 –140
RSV	Pneumoviridae	- ssRNA	enveloped	spherical	15	150 –250
Rhinovirus	Picornavirus	+ ssRNA	nonenveloped	icosahedral	7	15 – 30
MS2	Leviviridae	+ ssRNA	nonenveloped	icosahedral	3.6	23 – 28

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604 **Table 2: Isotherm parameters (K_F and n) and average relative error (ARE) of Freundlich**
 605 **models for the adsorption of SARS-CoV-2, RSV-A, RV-B, and MS2 in wastewater at 4°C**
 606 **and 22°C; SE, LE, and UE are the standard error, the lower SE bound, and the upper SE**
 607 **bound as reported by lm function in R, respectively. n and ARE are dimensionless.**

Lab-grown viruses	Virus structure	Temperature (°C)	K_F (LE-UE) (ml·g ⁻¹)	$n \pm SE$	ARE
SARS-CoV-2	Enveloped	4	18,000 (4,100-41,000)	0.81 ± 0.07	0.40
		22	270,000 (74,000-630,000)	0.64 ± 0.09	0.91
RSV-A	Enveloped	4	32,000 (2,000-67,000)	1.24 ± 0.02	0.25
		22	19,200 (10,000-60,000)	1.32 ± 0.21	1.18
RV-B	Nonenveloped	4	13,000 (1,500-28,000)	0.84 ± 0.03	0.15
		22	8,900 (2,400-21,000)	0.74 ± 0.07	0.39
MS2	Nonenveloped	4	18,000 (7,400-49,000)	0.66 ± 0.09	0.64
		22	2,000 (760-5,200)	0.70 ± 0.08	0.78

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611 **Table 3: Distribution coefficient ($K_d = C_s/C_w$) of endogenous SARS-CoV-2, RSV, RV, and F+**
612 **coliphage RNA in wastewater influent; SD is the standard deviation across triplicate**
613 **subsamples**

Wastewater treatment plant	$K_d \pm SD$ (ml·g ⁻¹)			
	SARS-CoV-2	RSV	RV	F+ coliphage
Gilroy	3,500 ± 820	1,300 ± 650	7,200 ± 1,400	490 ± 380
San Jose	11,000 ± 6,200	2,000 ± 760	16,000 ± 4,100	2,800 ± 1800
Sunnyvale	12,000 ± 7,600	1,800 ± 460	8,000 ± 4,800	950 ± 710
SVCW	5,600 ± 4,300	1,300 ± 450	11,000 ± 7,200	7,400 ± 6,300
SEP	4,400 ± 1,400	500 ± 320	2,100 ± 1,400	1,200 ± 760
OSP	3,000 ± 2,500	700 ± 470	2,800 ± 1,000	1,300 ± 470

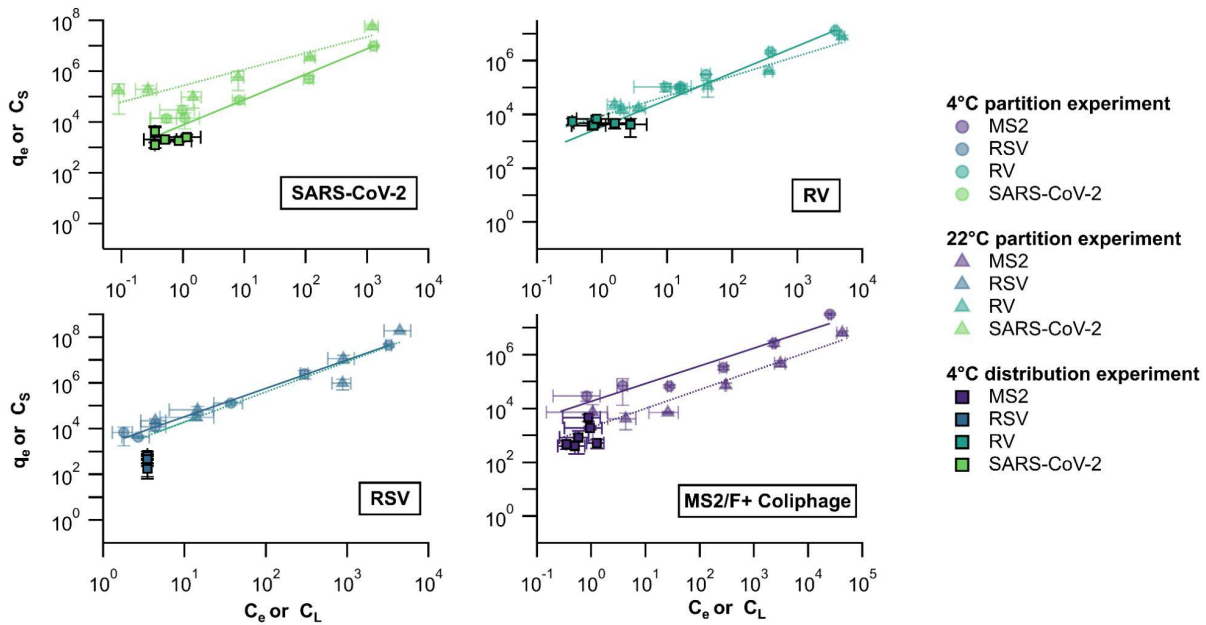
614

615 **Table 4: Results from previous experiments measuring the concentration of viral genetic**
 616 **markers in the solid and liquid fractions of wastewater matrices.**

Study	Virus	Wastewater Matrix	Sample processing	K_F or K_d ($\text{ml}\cdot\text{g}^{-1}$)
Graham et al. 2021 ¹⁵	SARS-CoV-2	Paired wastewater influent and primary sludge samples	Solid and liquid fractions were separated by centrifugation. The supernatant from influent samples was further processed using PEG precipitation method. Viral concentrations (N1 and N2) were measured using RT-ddPCR.	K_d : 350–3,100 Calculated/reported as the solid-influent concentration ratio.
Li et al. 2021 ¹⁶	SARS-CoV-2	Wastewater influent	Solid and liquid fractions were separated by centrifugation. Liquids were further processed using PEG precipitation method. Viral concentrations (N1, N2, and E gene) were measured using qPCR.	K_d : 4,000–20,000 Calculated/reported as the solid-liquid concentration ratio.
Kim et al. 2022 ³⁶	SARS-CoV-2	Paired wastewater influent and primary sludge samples	Solid and liquid fractions were separated using different processing techniques. Viral concentrations (N1 and N2) were measured using RT-ddPCR.	K_F : 1,000–100,000 Calculated using the Freundlich model; solid to influent ratio.
Kim et al. 2023 ³⁰	SARS-CoV-2	Wastewater collected from sewer network	Solid and liquid fractions were separated by centrifugation. Liquids were further processed using a 0.45 μm pore size filter. Viral concentrations (N and S gene) were measured using RT-ddPCR.	K_d : 8,600, 16,000 Calculated/reported as the solid-liquid concentration ratio
Wolfe et al. 2022 ³⁷	Influenza A	Paired wastewater influent and primary sludge samples	Solid and liquid fractions were separated by centrifugation. The supernatant from influent samples was further processed using PEG	K_d : 1,000 Calculated/reported as the solid-influent concentration ratio.

			precipitation method. Viral concentrations (N1 and N2) were measured using RT-ddPCR.	
Mercier et al. 2022 ²	Influenza A	Wastewater and primary sludge samples	Solid and liquid fractions were separated by centrifugation and filtration using a 0.45 μm pore size filter. Viral concentrations were measured in settled solids, suspended solids, and liquid fractions (supernatant) using RT-qPCR.	Authors only reported the percent of viral RNA adsorbed onto wastewater solids, K_d was not reported
Wolfe et al. 2023 ¹⁹	Mpox	Paired wastewater influent and primary sludge samples	Influent samples were processed using an affinity-based capture method with magnetic hydrogel Nanotrap Particles with Enhancement Reagent 1. Primary sludge samples were centrifuged to obtain solids. Viral concentrations were measured using RT-ddPCR.	K_d : 1,000 Calculated/reported as the solid-influent concentration ratio.
Yin et al. 2018 ²⁰	Adenovirus	Primary and secondary sludge	Solid and liquid fractions were separated by centrifugation. Viral concentrations were measured in the liquid fraction (supernatant) using qPCR. Viral concentrations in solids were estimated using a mass balance equation.	K_F : 37,000, 40,000 Calculated using the Freundlich model.
Yang et al. 2022 ¹¹	MS2, Phi6, Phix174, T4	Primary sludge	Solid and liquid fractions were separated by centrifugation and filtration using a 0.22 μm pore size filter. Viral concentrations were measured using qPCR.	K_F : 4.1×10^6 , 5.4×10^5 , 1.2×10^5 , and 8.5×10^3 for Phi6, MS2, T4, and Phix174, respectively.

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Figure 1: q_e and C_e from partition experiments at 4°C (circles) and 22°C (triangles) and C_s and C_L from distribution experiment at 4°C (squares, with black edges). Lines represent the Freundlich isotherm model and error bars represent the standard deviation across triplicate subsamples.